Microbial Production and Consumption of Greenhouse Gases:
Methane, Nitrogen Oxides, and Halomethanes
Edited by John E. Rogers and William B. Whitman
© 1991 American Society for Microbiology, Washington, DC 20005

Chapter 13

# Cycling of NO<sub>x</sub> in Tropical Forest Canopies

Daniel J. Jacob and Peter S. Bakwin

Nitrogen oxides ( $NO_x = NO + NO_2$ ) play a central role in the chemistry of the troposphere. They regulate the photochemical production of ozone and the abundance of the hydroxyl radical (OH), which is the main oxidant for a number of trace gases including methane. Sources of  $NO_x$  to the troposphere include fossil fuel combustion, biomass burning, biogenic emissions from soils, and lightning (Logan, 1983; Table 1). Atmospheric oxidation of  $NO_x$  takes place on a time scale on the order of 1 day and produces nitric acid which is removed by deposition; this deposition is a major source of nitrogen and acidity to terrestrial ecosystems.

Simulation of NO<sub>x</sub> is currently a top priority in the development of global models for atmospheric chemistry (Brost et al., 1988; Levy and Moxim, 1989). However, as shown in Table 1, the size of the biogenic source is a serious uncertainty. Biogenic emissions could dominate the budget of NO<sub>x</sub> over the continental tropics, where soil NO<sub>x</sub> fluxes are high and anthropogenic emissions are relatively low (Johansson et al., 1988; Johansson and Sanhueza, 1988; Jacob and Wofsy, 1988, 1990). Proper accounting of this source requires consideration not only of the magnitude of soil NO<sub>x</sub> emissions (Davidson, this volume), but also of the fraction of those emissions that is lost by deposition to vegetation during transport from the soil to canopy top. The latter effect has been so far neglected in the construction of atmospheric source inventories, but we will argue here that it can reduce considerably the export of NO<sub>x</sub> from forested canopies (by about a factor of 4). As a result, current global estimates of the biogenic source of NO<sub>x</sub> to the atmosphere may be seriously exaggerated.

## BIOSPHERE-ATMOSPHERE EXCHANGE OF NO.

The biosphere is both a source and a sink for atmospheric  $NO_x$ . Microbes in soils emit  $NO_x$  (mainly as NO), while vegetation scavenges  $NO_x$  (mainly as  $NO_2$ ). The mechanisms for  $NO_x$  deposition to vegetation have been reviewed recently by Johansson (1989). The principal route appears to be uptake of  $NO_2$  by the plant stomata, followed by reduction to nitrite at the mesophyll and assimilation via

Daniel J. Jacob and Peter S. Bakwin • Department of Earth and Planetary Sciences and Division of Applied Sciences, Harvard University, 29 Oxford Street, Cambridge, Massachusetts 02138.

Table 1. Global source inventory for NOx in the troposphere

Source	NO <sub>x</sub> (Tg of N year - 1)
Fossil fuel combustion	. 21 (14–28)
Biomass burning	. 12 (4–24)
Soil emissions	. 8 (4–16)
Lightning	8 (2–20)
Atmospheric oxidation of NH <sub>3</sub>	. 1–10
Oceans	. < 1
Stratospheric input	. < 1
Total	. 25–99

<sup>&</sup>quot; Data from Logan (1983). Numbers are best estimates and possible ranges

nitrite reductase (Rogers et al., 1979). Uptake of NO is negligibly slow compared to uptake of  $NO_2$ . Johansson (1989) stressed the need to consider the balance between soil emissions of NO on the one hand and  $NO_2$  deposition to vegetation and to the ground on the other hand in assessing the flux of  $NO_x$  between the biosphere and the atmosphere. Indeed, measurements in polluted regions have documented net downward fluxes of  $NO_x$  to vegetation, due to the uptake of  $NO_x$  advected from anthropogenic sources upwind (Delany and Davies, 1983; Delany et al., 1986).

We wish to examine how the biosphere-atmosphere exchange of  $NO_x$  over tropical forests should be treated in regional and global models for atmospheric chemistry. These models generally use the top of the canopy as the lower boundary. The biogenic source of  $NO_x$  in the model is defined by the ventilation flux,  $F_v$  (molecules per square centimeter per second), of biogenic  $NO_x$  at canopy top, viz.:

$$F_v = \alpha E_{NO} \tag{1}$$

where  $E_{NO}$  is the soil emission flux of NO (molecules per square centimeter per second) and  $\alpha$  is an "export efficiency" ( $0 \le \alpha \le 1$ ) representing the fraction of  $NO_x$  emitted by soil that is ventilated to the atmosphere above the canopy. Estimates for  $\alpha$  will be presented below. The ventilation flux,  $F_{tv}$ , is balanced by a downward flux,  $F_d$  (molecules per square centimeter per second), of  $NO_x$  from the atmosphere to the canopy, which is usually expressed in terms of a deposition velocity,  $V_d$  (centimeters per second):

$$F_d = V_d(NO_x) \tag{2}$$

where  $(NO_x)$  is the concentration of  $NO_x$  (molecules per cubic centimeter) immediately above the canopy. Estimates for  $V_d$  will also be presented below. The net upward flux of  $NO_x$  at canopy top (i.e., at the lower boundary of the model) is the difference  $F_v - F_d$ .

Rapid chemical cycling between NO and  $NO_2$  takes place inside a forest canopy (Fig. 1; Table 2). NO is oxidized to  $NO_2$  by  $O_3$  subsiding from aloft and by

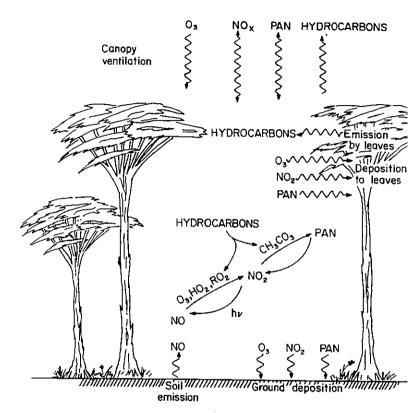


Figure 1. Chemical cycling of NO<sub>x</sub> in a forest canopy.

peroxy radicals produced within the canopy from decomposition of hydrocarbons (e.g., isoprene and terpenes). The oxidation of NO takes time scale of minutes, less than typical time scales for canopy ventilation (et al., 1990). In the daytime  $NO_2$  is photolyzed back to NO, and a equilibrium between NO and  $NO_2$  is established where the  $NO_2/NO$  mo of order unity. At night there is no mechanism for reaction of  $NO_2$  back that NO may be quantitatively converted to  $NO_2$ .

The  $NO_x$  budget inside the canopy is complicated by conversion  $CH_3C(O)OONO_2$  (peroxyacetylnitrate, abbreviated as PAN). This con driven by  $CH_3CO_3$  (peroxyacetyl) radicals generated locally from phot oxidation of biogenic hydrocarbons (Jacob and Wofsy, 1990). The lifetin against thermal decomposition back to  $NO_2$  is relatively short at the ten found in tropical forest canopies (e.g., 35 min at 300 K), so that a equilibrium may be established between  $NO_2$  and PAN. Dependin efficiency of PAN deposition to vegetation and to the ground, the equilibrium could affect the export of biogenic  $NO_x$  out of the canopy.

Table 2. Important reactions cycling NO<sub>x</sub> inside a forest canopy

Rate constant (cm $^3$ molecule $^{-1}$ s $^{-1}$ or s $^{-1}$ )	
$2.2 \times 10^{-12} e^{-1.430/T a}$	
$5.6 \times 10^{-36}$	
$4.3 \times 10^{-6b,c}$	
$2.5 \times 10^{-11} e^{409/T d}$	
$4.2 \times 10^{-12} e^{180/T}$	
Fast <sup>*</sup>	
$3.7 \times 10^{-12} e^{240/T}$	
$8.4 \times 10^{-4b.f}$	
$4.7 \times 10^{-12}$	
$1.95 \times 10^{16} e^{-13.543/T}$	

" T is temperature in units of K.

b Photolysis rate constant at noon at canopy top, computed at the equator with a Rayleigh scattering code assuming 30% opaque cloud cover overhead, an O<sub>3</sub> column of 7.1 × 10<sup>18</sup> molecules cm<sup>-2</sup>, and a canopy albedo of 0.1. The photolysis rate constants inside the canopy are reduced because of extinction of light by leaves; correction factors are given in the text.

<sup>c</sup> For  $P_{H_2O} = 24.7 \text{ mb}$ .

<sup>d</sup> Oxidation of hydrocarbons by OH produces organic peroxy radicals (RO<sub>2</sub>). The rate constant is strongly dependent on hydrocarbon type; the value shown here is for oxidation of isoprene.

Depending on the nature of the RO radical, products from this reaction may include HO<sub>2</sub>, organic peroxy

radicals, aldehydes, ketones, and dicarbonyls.

<sup>f</sup> Model calculations by Jacob and Wofsy (1990) indicate that this reaction (where methylglyoxal is produced from oxidation of isoprene) accounts for  $\approx 90\%$  of CH<sub>3</sub>CO<sub>3</sub> production in the Amazon forest canopy.

The data base of NO<sub>x</sub> flux measurements is limited, due in part to the difficulty of measuring NO<sub>x</sub> concentrations accurately (Johansson, 1989). In particular there are at this time no reliable data for NOx fluxes over forest canopies. Ideally, a field experiment designed to study the biosphere-atmosphere exchange of NO<sub>x</sub> over a forest should include the following measurements: (i) soil emission fluxes of NO, (ii) turbulent fluxes of  $NO_x$  at canopy top, (iii) concentrations of trace gases inside and just above the canopy (NO, NO<sub>2</sub>, O<sub>3</sub>, PAN, and isoprene are key species), (iv) leaf and ground resistances to deposition, and (v) rates of vertical mass exchange within the canopy and at canopy top. The data set collected during the NASA/ ABLE-2B expedition to the Amazon forest in the wet season of 1987 (Harriss et al., 1990) comes closest at this time to meeting the above requirements. As part of this expedition, extensive data were collected from a 40-m-high tower erected in a terra firme forest 20 km northeast of Manaus, Brazil. Trace gas concentrations at high altitudes were measured by aircraft. We will see below that these data afford important constraints on the export of biogenic NOx out of the Amazon forest canopy.

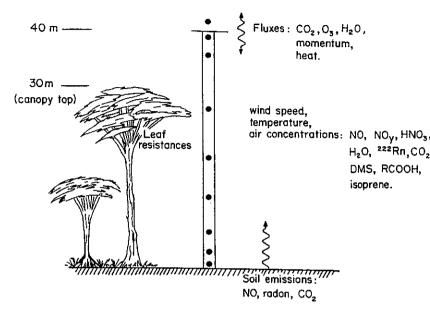


Figure 2. Ensemble of measurements made at the Ducke Forest Reserve, 20 km northed Manaus, Brazil, during the ABLE-2B expedition in April-May 1987. Fluxes and air contractions were measured from a 40-m-high tower erected through the forest canopy. I were measured at the top of the tower by the eddy correlation method. Air concentrations were measured at several altitudes, indicated by the solid circles. Soil emissions measured using static (closed) chambers. Leaf resistances were measured by Roberts (1990) during earlier expeditions at the site. Further details on investigators and methogiven by Harriss et al. (1990).

## OBSERVATIONAL CONSTRAINTS ON THE NO<sub>x</sub> BUDGET IN THE AMAZON FOREST CANOPY

### Overview of the ABLE-2B Data

Figure 2 gives an overview of the measurements available from the ABI tower site. The forest canopy was 30 m high, with emergents up to 35 m and area index of 7 (the leaf area index is the total leaf area per unit area of air colcounting only one side of leaf; it is a dimensionless measure of vegetation loac Soil fluxes of NO from the predominant clay soils averaged 8.9  $\times$  10° mole cm $^{-2}$ s $^{-1}$  with no significant diurnal variation (Bakwin et al., 1990b). Conce tions of NO inside the canopy decreased sharply with altitude (Fig. 3), refle the emission from soil and the rapid oxidation to NO<sub>2</sub> (Bakwin et al., 19 Maximum concentrations of NO were observed at night when canopy ventil was restricted and O<sub>3</sub> concentrations were near zero. Aircraft measurements a to 300 m altitude indicated NO concentrations of 12  $\pm$  7 parts per trillic volume (ppt), much lower than inside the canopy.

Concentrations of NO<sub>y</sub>, representing the sum of all reactive nitrogen o

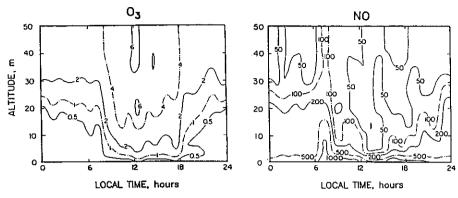
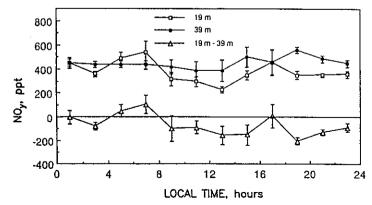


Figure 3. Average concentrations of  $O_3$  and NO measured at the ABLE-2B tower site as a function of time of day. Units are parts per billion by volume (ppb) for  $O_3$  and parts per trillion by volume (ppt) for NO.

(including NO<sub>x</sub>, HNO<sub>3</sub>, PAN, and other organic nitrates), were measured at 19 and 39 m altitude by Bakwin et al. (1990a) (Fig. 4). Values at 19 m averaged 400 ppt with little diurnal variation; values at 39 m were slightly higher, implying a net downward flux of NO<sub>y</sub> to the canopy. The principal components of NO<sub>y</sub> were not clearly identified. Concentrations of HNO<sub>3</sub> were only 20 to 50 ppt (Talbot et al., 1990), and concentrations of PAN measured from aircraft at 100 to 300 m altitude were less than 20 ppt (Singh et al., 1990). Concentrations of NO<sub>2</sub> were not measured. Model calculations (Jacob and Wofsy, 1990) suggest that NO<sub>2</sub> could have accounted for most of the NO<sub>y</sub> in the canopy at night, but not in the daytime.



**Figure 4.** Average concentrations of  $NO_y$  measured at the ABLE-2B tower site as a function of time of day at 19 and 39 m altitude. Figure adapted from Bakwin et al. (1990a).

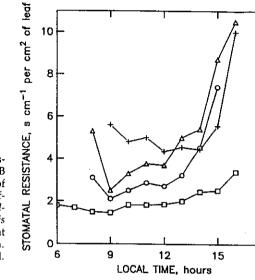


Figure 5. Average stomatal resistances measured at the ABLE-2B tower site as a function of time of day for four species present at different altitudes. Symbols: □, Piptadenia suaveolens at 33 m; ○, Naucleopsis glabra at 17 m; △, Gustavia angusta at 8 m; +, Scheelea sp. at 0 to 3 m. Figure adapted from Roberts et al. (1990).

Organic nitrates subsiding from aloft were probably the major contributor daytime NO<sub>v</sub>.

Concentrations of  $O_3$  within the canopy (Fig. 3) were very low compare values commonly observed in surface air (Logan, 1985), reflecting the redeposition of  $O_3$  to the forest vegetation. Deposition velocities of  $O_3$  measured a maltitude averaged 1.8 cm s<sup>-1</sup> in the daytime (Fan et al., 1990), a factor of 2 his than typical values observed over mid-latitudes forests (Lenschow et al., 1982) night the deposition velocities for  $O_3$  were lower but still substantial (averaging cm s<sup>-1</sup>).

Deposition of O<sub>3</sub>, NO<sub>2</sub>, and other gases to vegetation is thought to take primarily at the stomata of leaves (Hicks et al., 1987). Stomatal and bound (leaf-atmosphere) resistances to water vapor transfer were measured at the to site by Roberts et al. (1990) during expeditions in 1983 to 1985 which prece ABLE-2B. The boundary resistances were always small compared to the stom resistances. The stomatal resistances were themselves quite low (Fig. 5), consist with observations in other tropical forests (J. Roberts, personal communicat 1989), and presumably reflecting the high insolation and the availability of wallowest values for the stomatal resistances were observed for trees in the upcanopy; these increased gradually from mid-morning to mid-afternoon, possible a result of water stress.

The low stomatal resistances of the Amazon forest vegetation, together the high leaf area index, indicate the potential for rapid uptake of trace gases. can derive the stomatal resistances to uptake of a particular gas by using measured values for water vapor and scaling by the ratio of molecular diffusivi

Fan et al. (1990) explained their observed daytime deposition velocities for  $O_3$  during ABLE-2B by using a standard leaf resistance model (Hicks et al., 1987) constrained with the stomatal resistance data of Roberts et al. (1990). The deposition of  $O_3$  observed at night (when the stomata were closed) suggests that significant  $O_3$  uptake took place also at the outer (cuticular) surfaces of leaves (Fan et al., 1990).

A key variable regulating the export of biogenic NO, out of the canopy is the canopy ventilation rate. Much of the canopy ventilation at the ABLE-2B site occurred by episodic large-scale downdrafts flushing the entire canopy (Fitzjarrald and Moore, 1990; Fitzjarrald et al., 1990). This ventilation mechanism is typical of forest environments in general (Gao et al., 1989) and cannot be properly described with standard mixing-length models for atmospheric turbulence. Vertical mass exchange rates during ABLE-2B can nevertheless be constrained quite well by using the observed vertical distributions and fluxes of three chemical tracers: radon-222, CO<sub>2</sub>, and O<sub>3</sub> (Trumbore et al., 1990; Fan et al., 1990). Radon-222 is a particularly useful tracer as it is released at a relatively constant and uniform rate by the soil. and it is removed from the atmosphere solely by radioactive decay (half-life, 3.8 days). Trumbore et al. (1990) used the radon-222 and CO2 data from ABLE-2B to derive air residence times (7) within the 0-40-m column of 5.5 h at night and  $\leq 1$ h in the daytime. Sharp vertical gradients of radon-222 concentration were observed in the lowest 2 m above ground, but above that altitude the concentrations were relatively uniform.

### A Box Model for the Amazon Forest Canopy

The ABLE-2B data place some important constraints on the export of biogenic  $NO_x$  from the forest to the atmosphere. These constraints can be expressed in a simple way with a steady-state box model for the 0–40-m air column, including no explicit chemistry. In the next section we will present a more detailed process-based model which includes temporal and vertical resolution, as well as a full description of photochemistry; we will see that results from the box model capture to a good approximation the main features of the complicated model. The upper boundary of the box model is chosen at 40 m because the canopy ventilation times ( $\tau$ ) given by Trumbore et al. (1990) are defined with respect to that altitude. We define "biogenic  $NO_x$ " as the  $NO_x$  supplied to the 0–40-m column by soil emission, in contrast to "atmospheric  $NO_x$ " supplied to the column from aloft.

The ventilation flux  $(F_v)$  of biogenic  $NO_x$  through the top of the 0–40-m air column is given by:

$$F_{v} = \frac{\Delta Z}{\tau} (NO_{x}) \tag{3}$$

where (NO<sub>x</sub>) is the concentration in the column and  $\Delta Z = 40$  m. Replacing in equation 1, we obtain an expression for the export efficiency  $\alpha$  of biogenic NO<sub>x</sub> out of the canopy:

$$\alpha = \frac{(NO_x)}{E_{NO}} \frac{\Delta Z}{\tau}$$

where  $E_{NO}$  is the soil emission flux. Although (NO<sub>x</sub>) was not measure ABLE-2B, an upper limit is imposed by the NO<sub>y</sub> concentrations shown in Fi (NO<sub>y</sub>)  $\approx$  400 ppt implies that biogenic (NO<sub>x</sub>) < 400 ppt. At night when  $\tau \approx$  h, this upper limit on biogenic (NO<sub>x</sub>) places a severe constraint on  $\alpha$ . Inserting average observed value  $E_{NO} = 8.9 \times 10^9$  molecules cm  $^{-2}$  s  $^{-1}$  into equatic and converting (NO<sub>x</sub>) to units of molecules per cubic centimeter (1 ppt = 2.4  $10^7$  molecules cm  $^{-3}$ ), we find  $\alpha < 0.22$ . In the daytime, by contrast, the NO<sub>y</sub> afford no constraint on  $\alpha$ ; replacing  $\tau = 1$  h into equation 4 yields the useless r  $\alpha < 1.2$ . Most of the NO<sub>y</sub> in daytime was contributed by non-NO<sub>x</sub> specie opposed to nighttime (see Overview of the ABLE-2B Data, above). The upper I (NO<sub>x</sub>) < 400 ppt, is evidently not sufficiently stringent to constrain  $\alpha$  ir daytime.

An estimate for  $\alpha$  in the daytime can, however, be derived from the depos flux of biogenic  $NO_x$  to vegetation. Let R (seconds per centimeter) be the tota resistance to  $NO_2$  deposition per square centimeter of air column, and let (NO the concentration of biogenic  $NO_2$  in the 0–40-m column. The deposition flubiogenic  $NO_x$  is given by  $(NO_2)/R$  and represents the balance between the emission flux,  $E_{NO_x}$  and the ventilation flux,  $F_y$ :

$$\frac{(NO_2)}{R} = E_{NO} - F_v$$

By combining equations 3 through 5 we obtain:

$$\alpha = \frac{1}{1 + \frac{\tau (NO_2)}{R\Delta Z(NO_x)}}$$

The ratio  $(NO_2)/(NO_x)$  in daytime is of order 0.7 (see below). An upper limit I can be estimated by assuming that deposition of  $NO_2$  is restricted to the stomata. We then decompose R as the sum of a boundary resistance  $(R_b)$ , a stor resistance  $(R_s)$ , and a mesophyllic resistance  $(R_m)$  placed in series (Hicks e 1987):

$$R = \frac{R_b + \beta R_s + R_m}{L}$$

Here L=7 is the leaf area index,  $\beta \approx 1.6$  is the ratio of the molecular diffusiv of water vapor and NO<sub>2</sub>, and  $R_b$ ,  $R_s$ , and  $R_m$  are in units of seconds per centin per square centimeter of leaf. Assuming  $R_m=0$  (Wesely, 1989), mean value

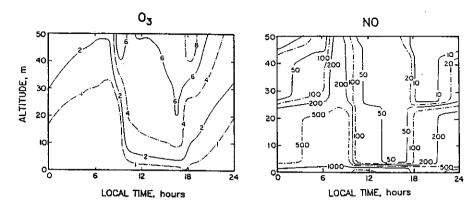


Figure 6. Concentrations of O<sub>3</sub> (ppb) and NO (ppt) simulated by the model of Jacob and Wofsy (1990), as a function of altitude and time of day. Results are in good agreement with observations (compare with Fig. 3). Figure adapted from Jacob and Wofsy (1990).

 $R_b$  and  $R_s$  of 0.5 and 3 s cm<sup>-1</sup> per cm<sup>2</sup> of leaf, respectively (Roberts et al., 1990), and  $\tau = 1$  h, we obtain  $\alpha = 0.56$ . Averaging the daytime and nighttime results yields a 24-h mean value of  $\alpha \le 0.39$ . It appears therefore that only a small fraction of NO emitted by soil is exported to the atmosphere above the canopy.

## ONE-DIMENSIONAL MODEL OF NO. CYCLING IN THE AMAZON FOREST CANOPY

The box model presented above suffers from obvious flaws, in particular the assumption of uniform  $\mathrm{NO_x}$  concentrations and the neglect of deposition to the ground and to the cuticles of leaves. An improved estimate of the  $\mathrm{NO_x}$  budget in the Amazon forest canopy can be obtained by using the one-dimensional process-based model of Jacob and Wofsy (1990). This model was developed to simulate the atmospheric chemistry observed during the ABLE-2B expedition; it has proven successful at reproducing various observations from tower and aircraft, in particular the NO and  $\mathrm{O_3}$  concentrations inside the canopy (Fig. 6). The full model extends from the ground to 2,000 m altitude and includes (i) vertical transport rates constrained by observations for radon-222,  $\mathrm{O_3}$ , and  $\mathrm{CO_2}$ ; (ii) detailed photochemistry describing the oxidation of biogenic hydrocarbons and the cycle of reactive nitrogen oxides; and (iii) biosphere-atmosphere exchange regulated by local vegetation density, leaf resistances, insolation, and temperature.

We consider here a truncated model domain, extending from 0 to 40 m altitude only, with fixed concentrations at 40 m as upper boundary conditions. The domain is subdivided into four grid cells (Table 3); cells 1 to 3 are in the canopy (0 to 30 m), and cell 4 is above the canopy (30 to 40 m). Each cell is assumed to be individually well mixed and exchanges air with adjacent cells by turbulent diffusion. The

Table 3. Structure of the Jacob and Wofsy (1990) model for the Amazon forest canopy<sup>a</sup>

Grid cell no."	Atmospheric column <sup>b</sup> (m)	Leaf area index <sup>c</sup>	Exchange velocity at grid cell top <sup>d</sup> (cm s <sup>- 1</sup> )	
			0900-1600	1600-0900
1	0-2	1	0.13	0.4
2	2-20	2	2	0.5
3	20-30	4	15	1
4	30-40	0	2	0.2

<sup>a</sup> The model is one-dimensional in the vertical and simulates concentrations and fluxes of chemically reactive trace gases in the air column between 0 and 40 m altitude. The top of the forest canopy is at 30 m altitude.

b The atmosphere extending from 0 m (ground level) to 40 m altitude is subdivided into four grid cells (1 through 4), which are assumed to be individually well mixed and to exchange mass with adjacent grid cells.

The leaf area index in each grid cell measures the local density of vegetation (square centimeters of leaf area per square centimeter of air column, counting only one side of the leaf). The sum of leaf area indices for all grid cells is 7 and represents the total leaf area index of the canopy.

<sup>d</sup> The exchange velocities define the rates of turbulent mass transfer between grid cells, as given by equation 8 in the text. We distinguish between a daytime regime (0900 to 1600 local time) when turbulence is vigorous and a nighttime regime (1600 to 0900 local time) when vertical motions are suppressed.

upward vertical flux,  $F_{i,j}$  (molecules per square centimeter per second), of spec i through the top of grid cell j is given by:

$$F_{i,j} = V_j \left[ \frac{n_{i,j}}{N_j} - \frac{n_{i,j+1}}{N_{j+1}} \right] \frac{N_j + N_{j+1}}{2}$$

where  $V_j$  (centimeters per second) is an exchange velocity,  $n_{i,j}$  (molecules per cu centimeter) is the concentration of species i in cell j, and  $N_j$  (molecules per cu centimeter) is the air density in cell j. The exchange velocities measure the intensof turbulence and are specified to fit the observed concentrations and vertical flu of radon-222,  $O_3$ , and  $CO_2$  (Table 3). The residence time in the 0–40-m model column for an inert tracer emitted at the ground is 0.6 h at noon and 5.0 h midnight, consistent with the radon-222 data of Trumbore et al. (1990).

Upper boundary concentrations for the model at 40 m are specified fr ABLE-2B observations. Very low concentrations of  $NO_x$  and PAN are imposed at m in order to suppress the downward flux of atmospheric  $NO_x$  and thus isolate from  $F_d$ . The simulations are iterated to a diurnal steady state, i.e., to a solut where concentrations show no net change over a 24-h cycle. The time scale to rethis steady state is less than 1 day, and hence we believe it to be a gapproximation of atmospheric conditions.

The total leaf area index of the canopy (L=7) is apportioned among the th canopy grid cells (0 to 30 m) as shown in Table 3. Half of the total leaf area is in upper canopy. Extinction of light by vegetation is computed by treating the cano

as a grey absorber and assuming a uniform angular distribution of leaves. The canopy optical depth at any altitude is then equal to half the leaf area index of the vegetation overhead (Verstraete, 1987). The resulting insolation near the ground at noon ( $\theta=13^{\circ}\text{C}$ ) is only 2.8% of the value at canopy top; the optical depth of the vegetation has thus a major effect on photochemical rates (in particular NO<sub>2</sub> photolysis).

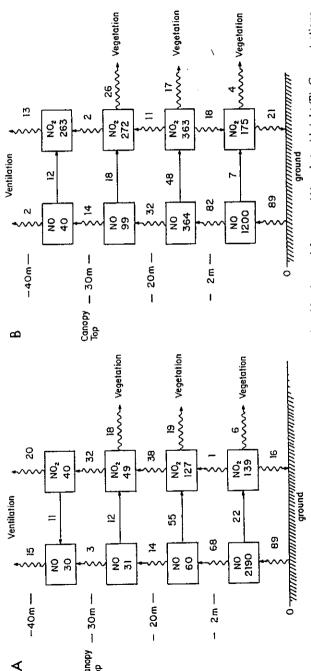
Natural hydrocarbons (mainly isoprene and acetaldehyde) are emitted by vegetation in each grid cell at a rate dependent on vegetation density, temperature, and insolation, as described by Jacob and Wofsy (1990). The hydrocarbon emission fluxes peak in the upper canopy at noon; 24-h average emission fluxes at canopy top are 9.1  $\times$  10<sup>10</sup> molecules cm<sup>-2</sup> s<sup>-1</sup> for isoprene and 4.0  $\times$  10<sup>9</sup> molecules cm<sup>-2</sup> s<sup>-1</sup> for acetaldehyde. Deposition of O<sub>3</sub>, NO<sub>2</sub>, and other reactive species is simulated in each grid cell using equation 7, with additional terms to describe deposition to the ground and to the leaf cuticles. For O<sub>3</sub> and NO<sub>2</sub> we assume  $R_m = 0$ , cuticular resistances of 10 s cm<sup>-1</sup> per cm<sup>2</sup> of leaf (Fan et al., 1990), and resistances to deposition at the ground of 2 s cm<sup>-1</sup> in the lowest grid cell (Johansson et al., 1988; Wesely, 1989). Values of  $R_b$  and  $R_s$  are taken from Roberts et al. (1990) and vary with altitude and time of day (see Fig. 5).

Figure 7 shows the main pathways for  $NO_x$  cycling within the 0–40-m air column at noon and at midnight. The  $NO_x$  budget reflects largely a balance between soil emissions of NO, deposition of  $NO_2$ , and ventilation. Conversion of  $NO_2$  to PAN turns out to be a minor process; the concentrations of PAN inside the canopy at noon are only  $\approx$  10% those of  $NO_2$ , and the upward flux of PAN at 40 m is only 4% that of  $NO_x$ . Production of PAN is inhibited by the optical thickness of the canopy, which suppresses the photochemical decomposition of biogenic hydrocarbons and hence the source of  $CH_3CO_3$  radicals.

The  $NO_x$  concentrations in the canopy are higher at night than in the daytime, partly because of the higher resistance to  $NO_2$  deposition and partly because of the lower canopy ventilation rate. At night NO is oxidized to  $NO_2$  by  $O_3$  and there is no conversion of  $NO_2$  back to NO, so that  $NO_x$  reaching the top of the canopy is  $\approx 90\%$   $NO_2$ . In the daytime, oxidation of NO is facilitated by the higher concentrations of  $O_3$  (Fig. 6) and by the presence of peroxy radicals produced from the decomposition of isoprene (which account for about half of total NO oxidation). Photolysis of  $NO_2$  maintains a daytime  $NO_2/NO_x$  ratio in the range 0.5 to 0.9 inside the canopy.

Inspection of Fig. 7 indicates that the uptake of  $NO_2$  by vegetation is roughly evenly distributed with altitude and that the ground accounts for 30% of total deposition. The export efficiency ( $\alpha$ ) of  $NO_{\infty}$  out of the canopy is 0.39 at noon and 0.17 at midnight; the 24-h average value is 0.25. The lower value of  $\alpha$  at night follows mainly from the restricted ventilation, which is only partially compensated by the higher resistance to deposition. The values of  $\alpha$  obtained here are somewhat lower than in the box model described above, principally due to the inclusion of ground deposition.

Deposition to the Amazon forest of atmospheric NO<sub>x</sub> supplied from aloft (equation 2) was investigated in a separate simulation in which we assumed zero



midnight (B). Concentrations unit area of air column Cycling of biogenic NO

soil emissions of NO and a fixed NO $_{\rm x}$  concentration at 40 m of 26 ppt (taken from Jacob and Wofsy [1990]). We found deposition velocities ( $V_d$ ) for NO $_{\rm x}$  at 40 m of 0.78 cm s  $^{-1}$  at noon and 0.13 cm s  $^{-1}$  at midnight; these values were insensitive to the speciation of NO $_{\rm x}$  assumed at 40 m because of the rapid chemical cycling between NO and NO $_{\rm 2}$ . Over 70% of total deposition of atmospheric NO $_{\rm x}$  took place in the upper canopy (20 to 30 m).

### **EXTENSION TO OTHER TROPICAL FOREST ENVIRONMENTS**

The fraction  $\alpha$  of biogenic  $NO_x$  emitted by soil that is ventilated to the atmosphere above the forest (export efficiency) depends on a number of environmental variables including (i) the magnitude of soil emission, (ii) the  $O_3$  concentration above the canopy, (iii) the leaf resistances to  $NO_2$  and  $O_3$  deposition, (iv) the canopy ventilation rate, and (v) the leaf area index. Figure 8 shows results from sensitivity simulations in which values for each of these variables were individually modified from the values used in the standard simulation described in the previous section. The range of perturbations was chosen to span the conditions likely to be encountered in tropical forests. The perturbations to R and  $\tau$  were applied by multiplying all leaf and ground resistances, or all exchange velocities, by a given factor relative to the values in the standard simulation.

An increase in the O<sub>3</sub> concentration above the canopy increases the NO<sub>2</sub>/NO<sub>2</sub> ratio and hence the scavenging of NO<sub>x</sub> by vegetation. At O<sub>3</sub> concentrations of >30 parts per billion by volume (ppb), the NO<sub>2</sub>/NO<sub>x</sub> ratio approaches unity through most of the canopy and a lower limit for  $\alpha$  is reached which is dependent solely on the relative rates of NO<sub>2</sub> deposition and ventilation. An increase in NO emission, by contrast, decreases the NO<sub>2</sub>/NO<sub>x</sub> ratio and hence increases α. High soil fluxes of NO could result theoretically in complete titration of O3 inside the canopy and consequently in very high values of  $\alpha$ . However, it seems unlikely that titration of O3 could ever occur since high soil fluxes of NO would foster photochemical production of O<sub>3</sub> in the atmosphere above the forest, thus increasing the supply of O3 to the canopy from aloft. Such a situation was indeed encountered during the dry season ABLE-2A expedition, which operated from the same site as ABLE-2B (Harriss et al., 1988). Soil emission fluxes of NO in the dry season averaged 5.2  $\, imes$ 1010 molecules cm 2 s 1, six times higher than in the wet season, but O3 concentrations averaged = 20 ppb at 40 m and 10 to 20 ppb inside the canopy (Kaplan et al., 1988). Model calculations for the dry season (Jacob and Wofsy, 1988) indicate that the concentrations of biogenic NO<sub>x</sub> above the canopy were sufficiently high to promote photochemical production of O<sub>3</sub>, explaining in part the relatively high O<sub>3</sub> levels (NO<sub>x</sub> from biomass burning was an additional explanation). From the average dry season values for  $E_{\rm NO}$  and  $O_{\rm 3}$  concentrations we derive a 24-h mean  $\alpha$  of 0.22, slightly lower than in the wet season (0.25).

Changes in the resistance to deposition, the canopy ventilation rate, or the leaf area index have remarkably little effect on  $\alpha$  because of negative feedbacks. For example, an increase in the resistance to deposition hinders the uptake of NO<sub>2</sub> but also of O<sub>3</sub>; it facilitates penetration of O<sub>3</sub> inside the canopy and hence increases the

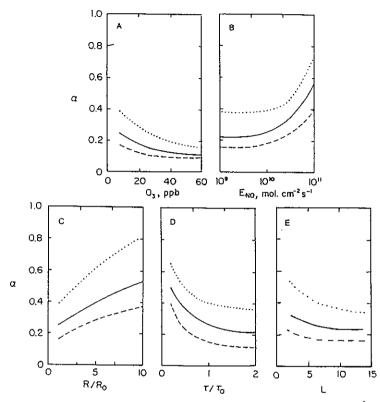


Figure 8. Export efficiency  $\alpha$  of biogenic  $NO_x$  out of the forest canopy as a function of environmental variables. The export efficiency represents the fraction of  $NO_x$  emitted by soil that is exported to the atmosphere above the canopy. The environmental variables include (A)  $O_3$  concentration at 40 m altitude, (B) soil emission flux of NO ( $E_{NO}$ ), (C) total resistance to deposition (R), (D) residence time of air inside the 0–40-m column ( $\tau$ ), and (E) canopy leaf area index (L). The perturbations to R and  $\tau$  are applied by multiplying the individual resistances and exchange velocities used in the standard simulation (corresponding to standard values  $R_0$  and  $\tau_0$ ) by a given factor. The figure shows 24-h average values for  $\alpha$  (solid line) and values at noon and midnight (dashed and dotted lines, respectively).

 $NO_2/NO_x$  ratio. Similar arguments can be made to explain the weak dependen of  $\alpha$  on canopy ventilation rate and on leaf area index.

#### CONCLUSIONS

Results from a process-based model for the Amazon forest atmosphere indic that only 25% of biogenic  $NO_x$  emitted by soil is exported to the atmosphere abothe canopy (35% in the daytime, 17% at night). The balance is deposited

vegetation during transport from the ground to canopy top. The deposited  $NO_{\times}$  may be assimilated into the plant material, thus recycling nitrogen within the ecosystem. The role of the canopy filter in limiting the export of biogenic  $NO_{\times}$  to the atmosphere appears to be important over a broad range of conditions encountered in tropical forest environments. Neglect of this effect in atmospheric chemistry models may lead to serious overpredictions of  $NO_{\times}$  concentrations above tropical continents.

Extension of our results to non-forest canopies is difficult because canopy ventilation rates and leaf surface resistances may be vastly different. High soil emissions of NO have been observed from wet savannas (Johansson and Sanhueza, 1988) and from fertilized cropland (Williams et al., 1988), but the ventilation of such open canopies could be much faster than for tropical forests, and the seasonality of leaf activity would need to be considered. Open chamber flux measurements by Slemr and Seiler (1984) indicate lower NO emissions from grassland than from adjacent bare soil, suggesting uptake of NO<sub>2</sub> by grass. However, Johansson and Granat (1984) observed the same effect when using a static chamber where O<sub>3</sub> (and hence NO<sub>2</sub>) should be depleted due to excess NO. More research is needed to explain these observations.

Acknowledgments. This work was funded by the National Science Foundation (NSF-ATM88-58974 and NSF-ATM-89-21119), the Host Foundation, and the Packard Foundation.

#### REFERENCES

- Bakwin, P. S., S. C. Wofsy, and S.-M. Fan. 1990a. Measurements of reactive nitrogen oxides (NO<sub>y</sub>) within and above a tropical forest canopy in the wet season. *J. Geophys. Res.* 95:16765–16772.
- Bakwin, P. S., S. C. Wofsy, S.-M. Fan, M. Keller, S. Trumbore, and J. M. da Costa. 1990b. Emission of nitric oxide (NO) from tropical forest soils and exchange of NO between the forest canopy and atmospheric boundary layers. J. Geophys. Res. 95:16755-16764.
- Brost, R. A., R. B. Chatfield, J. P. Greenberg, P. L. Haagenson, B. G. Heikes, S. Madronich,
   B. A. Ridley, and P. R. Zimmerman. 1988. Three-dimensional modeling of transport of chemical species from continents to the Atlantic Ocean. *Tellus* 40:358-379.
- Delany, A. C., and T. D. Davies. 1983. Dry deposition of NO<sub>x</sub> to grass in rural East Anglia. Atmos. Environ. 17:1391-1394.
- Delany, A. C., D. L. Fitzjarrald, D. H. Lenschow, R. Pearson, Jr., G. J. Wendel, and B. Woodruff. 1986. Direct measurements of nitrogen oxides and ozone fluxes over grasslands. J. Atmos. Chem. 4:429-444.
- Fan, S.-M., S. C. Wofsy, P. S. Bakwin, D. J. Jacob, and D. R. Fitzjarrald. 1990. Atmosphere-biosphere exchange of CO<sub>2</sub> and O<sub>3</sub> in the central Amazon forest. *J. Geophys. Res.* 95:16851–16864.
- Fitzjarrald, D. R., and K. E. Moore. 1990. Mechanisms of nocturnal exchange between the rain forest and the atmosphere. J. Geophys. Res. 95:16839–16850.
- Fitzjarrald, D. R., K. E. Moore, O. M. R. Cabral, J. Scolar, A. O. Manzi, and L. D. de Abreu Sa. 1990. Daytime turbulent exchange between the Amazon forest and the atmosphere. J. Geophys. Res. 95:16825-16838.
- Gao, W., R. H. Shaw, and K. T. Paw U. 1989. Observation of organized structure in turbulent flow within and above a forest canopy. *Boundary-Layer Meteorol*. 47:349–377.
- Harriss, R. C., et al. 1988. The Amazon Boundary Layer Experiment (ABLE 2A): dry season 1985. J. Geophys. Res. 93:1351-1360.
- Harriss, R. C., et al. 1990. The Amazon Boundary Layer Experiment: wet season 1987. J. Geophys. Res. 95:16721-16736.

- Hicks, B. B., D. D. Baldocchi, T. P. Meyers, D. R. Matt, and R. P. Hosker. 1987. A m preliminary resistance routine for deriving dry deposition velocities from me quantities. *Water Air Soil Pollut*. 36:311–330.
- Jacob, D. J., and S. C. Wofsy. 1988. Photochemistry of biogenic emissions over the A forest. J. Geophys. Res. 93:1477-1486.
- Jacob, D. J., and S. C. Wofsy. 1990. Budgets of reactive nitrogen, hydrocarbons, and over the Amazon forest during the wet season. J. Geophys. Res. 95:16737–16754.
- Johansson, C. 1989. Fluxes of NO, above soil and vegetation, p. 229-246. In M. O. A and D. S. Schimel (ed.), Exchange of Trace Gases between Terrestrial Ecosystems a Atmosphere. John Wiley & Sons, Inc., New York.
- Johansson, C., and L. Granat. 1984. Emission of nitric oxide from arable land. *Tellus* 36 Johansson, C., H. Rodhe, and E. Sanhueza. 1988. Emission of NO in a tropical savan a cloud forest during the dry season. *J. Geophys. Res.* 93:7180-7192.
- Johansson, C., and E. Sanhueza. 1988. Emission of NO from savanna soils during season. J. Geophys. Res. 93:14193-14198.
- Kaplan, W. A., S. C. Wofsy, M. Keller, and J. M. Da Costa. 1988. Emission of N deposition of O<sub>3</sub> in a tropical forest system. J. Geophys. Res. 93:1389–1395.
- Lenschow, D. H., R. Pearson, Jr., and B. B. Stankov. 1982. Measurements of ozone flux to ocean and forest. J. Geophys. Res. 87:8833-8837.
- Levy, H., II, and W. J. Moxim. 1989. Simulated global distribution and deposition of r nitrogen emitted by fossil fuel combustion. Tellus 41:256-271.
- Logan, J. A. 1983. Nitrogen oxides in the troposphere: global and regional budgets. J. G Res. 88:10785–10807.
- Logan, J. A. 1985. Tropospheric ozone: seasonal behavior, trends, and anthrop influence. J. Geophys. Res. 90:10463–10482.
- Roberts, J., O. M. R. Cabral, and L. F. De Aguiar. 1990. Stomatal and boundar conductances in an Amazonian Terra Firme rain forest. J. Appl. Ecol. 27:336–353.
- Rogers, H. H., J. C. Campbell, and R. J. Volk. 1979. Nitrogen-15 dioxide uptal incorporation by *Phaseolus vulgaris* (L.). Science 206:333-335.
- Singh, H. B., D. Herlth, D. O'Hara, L. Salas, A. L. Torres, G. L. Gregory, G. W. Sach: J. F. Kasting. 1990. Atmospheric peroxyacetyl nitrate measurements over the Br Amazon Basin during the wet season: relationships with nitrogen oxides and oz Geophys. Res. 95:16945-16954.
- Slemr, F., and W. Seiler. 1984. Field measurements of NO and NO<sub>2</sub> emissions from fer and unfertilized soils. *J. Atmos. Chem.* 2:1–24.
- Talbot, R. W., R. C. Harriss, M. O. Andreae, H. Berresheim, P. Artaxo, M. Garstang Harriss, K. M. Beecher, and S.-M. Li. 1990. Aerosol chemistry during the wet see central Amazonia: the influence of long-range transport. J. Geophys. Res. 95:16955-1
- Trumbore, S. E., M. Keller, S. C. Wofsy, and J. M. da Costa. 1990. Measurements of s canopy exchange rates in the Amazon rain forest using <sup>222</sup>Rn. J. Geophys. Res. 95: 16874.
- Verstraete, M. M. 1987. Radiation transfer in plant canopies: transmission of direct radiation and the role of leaf orientation. J. Geophys. Res. 92:10985–10995.
- Wesely, M. L. 1989. Parameterization of surface resistance to gaseous dry deposiregional-scale numerical models. *Atmos. Environ.* 23:1293–1304.
- Williams, E. J., D. T. Parrish, M. P. Buhr, and F. C. Fehsenfeld. 1988. Measurements NO<sub>x</sub> emissions in central Pennsylvania. J. Geophys. Res. 93:9539-9546.